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Floc Roll-up and its Implications for the Spacing of Inclined Settling Devices

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6 1 **Floc Roll-up and its Implications for the**
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9 2 **Spacing of Inclined Settling Devices**

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28 8 ***Key words:*** sedimentation, inclined sedimentation, plate settler, tube settler, capture
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30 9 velocity, floc roll-up, velocity gradient, fractal dimension

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33 10 ***Running title:*** FLOC ROLL-UP IN INCLINED SEDIMENTATION

Abstract

Inclined plate and tube settlers are commonly used to make sedimentation tanks more compact. Conventional design equations for inclined settling devices are based on obtaining a desired particle capture velocity, and these equations suggest that a suitable capture velocity can be achieved by reducing plate settler spacing or tube settler diameter below that specified in conventional design guidelines. Smaller spacing would reduce capital cost by decreasing sedimentation tank volume. However, the existing literature does not explain why smaller values of plate or tube spacing cannot be used, and the failure mechanism that sets the minimum spacing recommended in design practice has not been documented. This research shows that the fluid velocity gradient at the tube or plate surface is the limiting constraint for spacing, and for very small spacing, particles that settle on the solid surface are carried up the incline. This failure mode, termed “floc roll-up,” occurs when the terminal velocity of a floc along the incline is less than the upward velocity of the fluid at the center of the floc. This paper presents an experimental and theoretical investigation of the floc roll-up failure mode and its implications for plate settler design. A model is developed to explain the physics of floc roll-up, and experiments in a bench-scale water treatment apparatus demonstrate its effects on sedimentation performance.

Key words: sedimentation, inclined sedimentation, plate settler, tube settler, capture velocity, floc roll-up, velocity gradient, fractal dimension

INTRODUCTION

Inclined plates or tubes are often employed to reduce the required area of new sedimentation tanks or to retrofit existing tanks and increase their capacity (Reynolds and Richards, 1996). Inclined settling devices are widely used in water and wastewater treatment, and as such, there is a large existing body of research on the optimal configurations for these systems. Reducing the spacing of inclined settling devices would decrease the cost of conventional municipal-scale treatment facilities by allowing for shallower sedimentation tanks. The possible benefit of reduced spacing is particularly important when considering both the large requirement for water infrastructure upgrades across the industrialized world, and the need for affordable water treatment technologies to expand access to safe drinking water in the developing world. Unfortunately, the spacing of the inclined plates or tubes is one design parameter that has not been extensively studied.

Existing documentation reveals suggested spacing ranging over more than an order of magnitude - from 1.3 to 30 cm - based on empirical evidence (AWWA, 1999; Hansen and Culp, 1967). However, little to no theoretical basis exists for these guidelines, and the technical literature provides no exploration of how local flow conditions between the inclined plate or tube surfaces affect particle capture. An American Water Works Association design book recommends a spacing of 30 cm for vertical-flow sedimentation tanks with floc blankets and asserts that widely spaced plate settlers are more cost-effective in floc blanket clarifiers (AWWA, 1999); however, no detailed analysis is presented. In contrast, Ziolo (1996) recommends a 5 cm spacing for “ordinary” performance in general sedimentation applications and suggests that spacing could be adjusted based on the influent solids concentration. Experiments by Hansen and Culp (1967) showed that circular tube settlers could achieve up to a 96% reduction in turbidity with a spacing of 1.3 cm, a significantly smaller value than the 5 cm suggested by Ziolo.

In addition to the suggested spacing of inclined settling devices, existing literature also disagrees on the physical basis upon which this spacing should be determined. Some re-

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4 57 searchers have characterized inclined sedimentation performance in terms of the ratio of
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6 58 plate length, L , to center-to-center spacing D , where improved performance was correlated
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8 59 with higher L/D ratios. For example, Yao (1970) reports that good performance was achiev-
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10 60 able for L/D ratios between 20 and 40. Beyond a ratio of 40, increasing the length of the
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12 61 plates or tubes yielded diminishing returns in terms of particle removal performance. Willis
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14 62 (1978) indicates that the greatest concern for small plate spacing is the danger of high ve-
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16 63 locities sweeping settling sludge out with finished water. Willis addresses this concern by
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18 64 designating the maximum tube loading rate as 1.7 mm/s and states that this tube settler
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20 65 design velocity “has proven reasonably satisfactory” in field applications.

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22 66 In the research presented here, laboratory studies and model calculations are used to
23
24 67 explore the quantitative relationship between flow conditions, geometry, and performance of
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26 68 inclined settling devices. A failure mode termed “floc roll-up” is proposed as the fundamental
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28 69 physical constraint on reduced spacing in inclined sedimentation systems. The theoretical
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30 70 effects of floc roll-up are explored and presented alongside experimental studies with a con-
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32 71 trolled bench-scale system.

33 34 72 35 36 37 73 **THEORETICAL MODEL DEVELOPMENT**

38 39 40 74 *Capture velocity and the floc roll-up condition*

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42 75 Successful capture of a floc by an inclined settling device requires three steps. First,
43
44 76 the floc must be able to settle onto the surface of a plate or tube settler; second, it must
45
46 77 slide down the incline to reach the lower section of the sedimentation tank; and third, the
47
48 78 floc must be removed from the lower section of the sedimentation tank. The settle step is
49
50 79 well characterized and is based on the geometry and flow through the inclined settlers. The
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52 80 second step, or slide step, is the focus of this paper.

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54 81 A schematic of an inclined sedimentation device is shown in Figure 1. For such a device -
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56 82 i.e., a tube settler with ends perpendicular to the axis - the settle capture velocity (referred
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4 83 to as critical velocity in some of the previous literature) is given by Equation (1) (Schulz
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6 84 and Okun, 1984):

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$$V_{Settle} = \frac{V_{\alpha}}{\frac{L}{D} \cos \alpha + \sin \alpha} \quad (1)$$

11
12 86 where L is the length of the tube settler, D is the tube diameter, α is the angle of inclination,
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14 87 V_{α} is the average fluid velocity in the tubes, and V_{Settle} is the terminal velocity of the slowest-
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16 88 settling particle that is reliably captured. Equation (1) suggests that tube settler performance
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18 89 (as manifest by V_{Settle}) is maintained as long as the ratio of L/D is constant for a fixed V_{α} .
19
20 90 Thus, according to this theory, it should be possible to reduce the required length of the
21
22 91 tube settlers by decreasing their diameter, and it is not clear in Equation (1) why the value
23
24 92 of D should not simply be made as small as possible.

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27 93 “Floc roll-up” is presented here as a failure mode that places a lower limit on the allow-
28
29 94 able diameter of tube settlers or the allowable spacing of plate settlers. After a floc settles
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31 95 on the lower surface of an inclined settling device, it continues to experience an upward
32
33 96 drag caused by the fluid flow, and the velocity at the centerline of the floc increases if the
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35 97 spacing between the plates or the diameter of the tubes is decreased. This is the case even
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37 98 if a constant average fluid velocity is maintained, because the velocity gradient at the wall
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39 99 increases as the spacing decreases. The gravitational force acts in the direction of down the
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41 100 incline, while fluid drag acts in the direction of up the incline. When the fluid drag and grav-
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43 101 itational forces balance, the floc remains stationary. This force balance sets the minimum
44
45 102 required spacing for plate or tube settlers. The force balance can also be obtained equating
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47 103 the velocity caused by fluid drag at the center line of the floc to the floc’s terminal settling
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49 104 velocity (in the absence of fluid flow) along the slope.

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106 *Fluid velocity at the floc centerline*

107 The velocity of the fluid at the centerline of a floc can be obtained from the parabolic
108 velocity profile in fully-established laminar flow. This velocity profile $v_z \{r\}$ can be found by
109 solving the Navier-Stokes equations for laminar conditions (Munson et al., 1999), as shown
110 in Equation (2):

$$111 \quad v_z \{r\} = \frac{8V_\alpha}{D^2} \left(\frac{D^2}{4} - r^2 \right) \quad (2)$$

112 where r is a coordinate normal to the tube axis and is set to zero at the middle of the circular
113 tube. The equation for the velocity gradient evaluated at the tube wall is:

$$114 \quad \left. \frac{dv_\alpha}{dr} \right|_{D/2} = \frac{8V_\alpha}{D} \quad (3)$$

115 For flow between plates separated by a distance D , the laminar flow parabolic velocity profile
116 is described by Equation (4) (Munson et al., 1999):

$$117 \quad v_z \{r\} = \frac{6V_\alpha}{D^2} \left(\frac{D^2}{4} - r^2 \right) \quad (4)$$

118 In the following discussion a tubular geometry is assumed. However, a comparison of Equa-
119 tions (2) and (4) shows that differences in geometry can be accounted for by a constant
120 - i.e., a factor of 6/8 is needed to convert from tubes to plates. While other geometries
121 such as hexagonal tubes are also used in inclined settling devices, tubes and plates represent
122 extremes and the following discussion can therefore be generalized to other configurations.

123 Inclined settlers are typically designed based on the vertical component of fluid velocity.
124 The relationship between the velocity in the direction of flow (i.e. along the incline) and the
125 vertical component is:

$$126 \quad V_\uparrow = V_\alpha \sin \alpha \quad (5)$$

127 where V_\uparrow is the average velocity in the vertical direction. An approximate equation for the
128 fluid velocity near the tube wall and parallel to the incline as a function of distance from the

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4 129 wall can be obtained by linearizing the velocity gradient at the wall of the tube, using the
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6 130 derivative in Equation (3). This, in turn, allows for the velocity v_α experienced at the center
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8 131 of a floc of diameter d_{Floc} resting on the wall of a circular tube to be approximated by:

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$$v_\alpha \left\{ \frac{D}{2} - \frac{d_{Floc}}{2} \right\} \approx \frac{8V_\alpha}{D} \left(\frac{d_{Floc}}{2} \right) = \frac{4V_\alpha d_{Floc}}{D} \quad (6)$$

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15 133 Note that this is a valid approximation when the floc diameter d_{Floc} is very small compared
16
17 134 to the diameter D of the tube settler.
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22 136 *Floc terminal settling velocity along the inclined surface*
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24 137 The terminal velocity of flocs sliding down the surface of an incline in the absence of fluid
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26 138 flow depends on their porosity, density, and diameter. Floc density can be approximated
27
28 139 based on a fractal model (Weber-Shirk and Lion, 2010). Adachi and Tanaka (1997) model
29
30 140 the terminal velocity V_T of a vertically-settling floc using Equation (7):
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33 141
$$V_T = \left(\frac{gd_{Floc_0}}{18\Phi\nu_{Water}} \right) \left(\frac{\rho_{Floc_0} - \rho_{Water}}{\rho_{Water}} \right) \left(\frac{d_{Floc}}{d_{Floc_0}} \right)^{D_{Fractal}-1} \quad (7)$$

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37 142 where d_{Floc_0} is the diameter of the primary colloidal particles, d_{Floc} is the floc diameter, Φ is a
38
39 143 floc shape factor, ν_{Water} is the kinematic viscosity of water, $D_{Fractal}$ is the fractal dimension,
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41 144 ρ_{Floc_0} is the density of the primary colloidal particle, and ρ_{Water} is the density of water. For
42
43 145 clay-aluminum flocs, Tambo and Watanabe (1979) report a drag coefficient equal to $45/Re$,
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45 146 where Re is the Reynolds number. Because the drag coefficient for spheres is equal to $24/Re$
46
47 147 for laminar conditions, the value of the floc shape factor Φ is taken to be $45/24$.
48

49 148 The fractal dimension is a critical parameter in describing floc behavior. Meakin (1987)
50
51 149 determined that flocs with two contact points have a fractal dimension of 2.13 and flocs
52
53 150 with three contact points have a fractal dimension of 2.19. Lambert et al. (2003) found that
54
55 151 the fractal dimension of *E. coli* flocs ranged from 1.90 to 2.20. Jarvis et al. (2005) report
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57 152 values of the fractal dimension ranging from 1.66 to 2.56 for coagulated natural organic
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4 153 matter. Nan et al. (2009) indicate that the fractal dimension of optimally-coagulated clay
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6 154 suspensions is 2.2. Li and Ganczarczyk (1989) analyzed the fractal dimensions of aggregates
7
8 155 based on settling test data and size-density relationships and determined an average fractal
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10 156 dimension of 2.3 for alum aggregates. The range of the floc fractal dimension is potentially
11
12 157 quite large.

13
14 158 The terminal settling velocity of a floc sliding down a frictionless inclined surface in a
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16 159 quiescent fluid is obtained by using the component of gravitational force along the incline.
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18 160 The terminal velocity for laminar conditions is linearly proportional to the acceleration due
19
20 161 to gravity. The terminal velocity $V_{T\alpha}$ down the incline of the tube settler is therefore:

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$$V_{T\alpha} = V_T \sin \alpha \quad (8)$$

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27 163 where α is the angle between the horizontal and the inclined tube and V_T is the terminal
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29 164 settling velocity in the vertical direction.

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34 165 *Floc roll-up predictions*

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36 166 Setting the fluid velocity at the floc centerline equal to the terminal sliding velocity of
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38 168 the floc along the slope gives the critical condition for floc roll-up, as shown in Equation (9):

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$$v_\alpha \left\{ \frac{D}{2} - \frac{d_{Floc}}{2} \right\} = V_{T\alpha} \quad (9)$$

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44
45 170 The relationship in Equation (9) is an approximation of the interaction of fluid drag and the
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47 171 gravitational force on the floc that will cause the floc to remain stationary on the incline.
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49 172 Substituting Equations (6), (7), and (8) into Equation (9) and solving for the terminal
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51 173 velocity V_T that will balance the velocity up the incline at the center of the floc yields a
52
53 174 general floc roll-up model:

$$V_{Slide} \approx V_{\uparrow} \left(\frac{4d_{Floc0}}{D \sin^2 \alpha} \right)^{\frac{D_{Fractal}-1}{D_{Fractal}-2}} \left[\frac{18V_{\uparrow} \Phi \rho_{Water} \nu_{Water}}{gd_{Floc0}^2 (\rho_{Floc0} - \rho_{Water})} \right]^{\frac{1}{D_{Fractal}-2}} \quad (10)$$

The sedimentation velocity V_{Slide} in Equation (10) is the terminal sedimentation velocity of the slowest-settling floc that can slide down the wall of an inclined settling device under the given conditions. Flocs with this terminal velocity (the slide velocity) will be held stationary on the incline because of a balance between gravitational forces and fluid drag. Flocs with a terminal velocity lower than V_{Slide} will be carried out the top of the tube (i.e., “roll up”) even if they settle onto the tube wall. Thus, the slide terminal velocity represents a constraint on the ability of tube settlers to capture flocs. Unlike the settle capture velocity in Equation (1), which is exclusively a property of the geometry and flow characteristics of the sedimentation tank, the slide capture velocity is a property of the floc as well as the sedimentation tank geometry and flow characteristics. This complexity is a result of the interaction between the size of the floc, its density, and the velocity gradient in the fluid.

The slide capture velocity V_{Slide} is plotted in Figure 2 as a function of upflow velocity for inclined tube settlers ($\alpha = 60^\circ$) of three different diameters, using the model in Equation (10) and assuming typical properties of clay-aluminum hydroxide flocs. Typical settle capture velocities ranging from 0.24 to 0.47 mm/s for flocculated water (Reynolds and Richards, 1996) are indicated in the shaded region of the figure. Inclined settling devices are designed to achieve a certain settle capture velocity, and floc roll-up will become a problem when V_{Slide} from Equation (10) is greater than this settle capture velocity. The floc roll-up failure mode provides a rationale for limiting the spacing or velocity in inclined settling devices. However, Figure 2 indicates that even small diameter tubes can have an average vertical fluid velocity greater than the 1.70 mm/s recommended by Hansen and Culp (1967) and still achieve a good settle capture velocity.

199 *Model limitations*

200 The model results presented above are strongly dependent on the assumed floc properties,
201 and the fractal flocculation model indicates that floc properties are dependent on the density
202 and size of the primary particle as well as the floc fractal dimension. Water treatment
203 plants must successfully remove a wide range of particles including clay, organic matter, and
204 coagulant. It is necessary to consider the worst-case scenario for floc roll-up, and further
205 research will be required to determine the properties of the flocs that are most difficult to
206 capture. Flocs formed from organic matter are expected to be less dense and thus more
207 susceptible to floc roll-up.

208 The slide capture velocity described above is defined based on a velocity profile for fully-
209 developed laminar flow. The length of the entrance region is a function of the Reynolds
210 number. The slide velocity model presented above assumes that the plate or tube settler is
211 sufficiently long to achieve fully-developed flow. In the entrance region of a tube or plate, the
212 velocity gradient at the wall is higher than in the fully-developed region. This means that it
213 is also important to consider the fate of flocs that are able to slide down from the region of
214 fully developed flow into the entrance region. As the velocity gradient at the wall increases,
215 some flocs sliding down from above will be unable to continue down the slope. As more
216 flocs slide down from above, the trapped flocs will accumulate and eventually become able
217 to progress further down toward the entrance region in a small avalanche. This avalanche
218 behavior is not part of the model, but it does not seem to present a constraint on the capture
219 of flocs.

221 **MATERIALS AND METHODS**

222 *Raw water preparation*

223 Influent water of approximately constant characteristics was prepared for the bench-scale
224 experiments using temperature-controlled, aerated tap water dosed with concentrated kaolin
225 clay. A schematic of the apparatus used to condition the raw water is shown in Figure

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4 226 3. Cornell University tap water from the university water filtration plant was used for all
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6 227 experiments. Typical properties of this water are: $\text{pH} = 7.5 \pm 0.3$; total hardness of 150
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8 228 mg/L as CaCO_3 ; total alkalinity of approximately 110 mg/L as CaCO_3 ; and total organic
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10 229 carbon of 2.0 mg/L (Foote et al., 2012).

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12 230 The water temperature was kept at 21°C using an electronic thermistor and computer-
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14 231 controlled addition of hot or cold water. This constant-temperature water was aerated to
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16 232 reduce the concentration of supersaturated gases. Kaolin clay was added to provide a con-
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18 233 trolled level of turbidity, and a concentrated stock of this clay was dosed into the raw water
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20 234 tank via a computer-controlled pinch valve. The raw water turbidity was continuously sam-
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22 235 pled and measured with a feedback control loop. This influent turbidity was set to 100 NTU
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24 236 with a coefficient of variation of $\pm 5\%$. Both the hot/cold solenoid valves and the clay stock
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26 237 pinch valve were automated with process control software written in LabVIEW, as described
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28 238 by Weber-Shirk (2009).

29 30 239 31 32 240 *Bench-scale treatment apparatus*

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35 241 A bench-scale experimental process train was used to treat the conditioned raw water via
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37 242 alum dosing, rapid mix, hydraulic flocculation, and upflow sedimentation. The sedimentation
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39 243 process included a floc blanket and an inclined tube settler. A diagram of the bench-scale
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41 244 treatment system is shown in Figure 4. Flow rates in the system were set by computer-
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43 245 controlled peristaltic pumps (Cole-Parmer, Vernon Hills, IL).

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45 246 A laboratory-grade alum solution was prepared daily with distilled water. Initial exper-
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47 247 iments showed that an alum dose of 45 mg/L (4.23 mg/L as Al) was suitable to produce a
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49 248 majority of particles with settling velocities greater than the settle capture velocity of the
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51 249 tube settlers. A raw water flow of 11.9 mL/s was used in all experiments. Raw water and
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53 250 alum were mixed by flowing through a 1 m length of 4.8 mm ID coiled tube with an energy
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55 251 dissipation rate of 0.1 W/kg .

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57 252 A tubular flocculator was used to prior to sedimentation to facilitate particle aggregation.
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The flocculator had a length L of 26 m, a coil diameter of 13.5 cm, an inner diameter D of 0.95 cm, a head loss h_L of 0.159 m, and a hydraulic residence time θ of 155 s. The Reynolds number for the tube flocculator was 1590, ensuring that laminar flow was maintained. For laminar conditions, the collision potential of a flocculator is proportional to $G\theta\phi_{Floc0}^{2/3}$ where G is the average velocity gradient in the flocculator, θ is hydraulic residence time, and ϕ_{Floc0} is the volume fraction of the primary particles, which is itself proportional to the concentration of clay and alum present in the system (Weber-Shirk and Lion, 2010). In these experiments, ϕ_{Floc0} was equal to 4.3×10^{-5} . The velocity gradient G in Equation (11) was based on the measured head loss through the flocculator:

$$G = \sqrt{\frac{gh_L}{\theta\nu_{Water}}} \quad (11)$$

The hydraulic residence time is based on the flow rate and geometry of the flocculator:

$$\theta = \frac{\pi D^2 L}{4Q} \quad (12)$$

Equations (11) and (12) were combined to calculate the value of $G\theta$.

$$G\theta = \frac{D}{2} \sqrt{\frac{\pi gh_L L}{Q\nu_{Water}}} \quad (13)$$

The value of $G\theta$ for the tube flocculator in this study was 15500. Although this value is on the low end of the range commonly used for hydraulic flocculators, it performed well in these experiments because the floc volume fraction was relatively high with 100 NTU raw water.

In a manner comparable to the experimental system described by Hurst et al. (2010), the sedimentation process consisted of an upflow clarifier with a floc blanket. The sedimentation tank upflow velocity was 1.2 mm/s and was set to be close to the optimal upflow velocity for floc blanket turbidity removal determined by Hurst et al. (2010). The height of the floc blanket was controlled by a floc wasting pump, which continuously pumped solids from

the floc blanket at a height 15 cm below the water level at the top of the column. All experiments presented in this paper were performed with a floc blanket. From the top of the sedimentation column, water flowed to one of three possible paths:

1. Some water was pulled through a tube settler inclined at $\alpha = 60^\circ$ at a rate set by the tube settler pump, making the tube settler velocity independent of the total system flow rate. The turbidity of this water was measured downstream of the tube settler.
2. Some water was drawn off the top of the sedimentation column to sample the turbidity above the floc blanket. This allowed the tube settler influent water to be tested so that the performance of the tube settler could be quantified.
3. The remainder of the flow passed over an overflow weir and was discharged.

Data acquisition and analysis

Turbidity readings were recorded at 5 s intervals using Micro TOL in-line turbidimeters (HF Scientific Model 20053, Ft. Myers, Florida) from the raw water tank; the floc blanket clarified effluent (i.e. the suspension within the sedimentation tank above the floc blanket); and the tube settler effluent. For experiments with small tube settler flow rates, a reservoir was used to accumulate tube settler effluent for sampling. Flow accumulated in a mixed container and was intermittently pumped through the tube settler effluent turbidimeter at 50 mL/min. Both mixing of the sample reservoir and the high sampling flow rate were employed to prevent settling of particles in the reservoir or the turbidimeter vial. Data was logged with the LabVIEW process controller software that also controlled the solenoid valves, pinch valves, and peristaltic pumps in the bench-scale apparatus.

Particle removal is reported below in terms of negative log fraction remaining, pC^* . This parameter is often called “log removal” and is defined by Equation (14):

$$pC^* = -\log \left(\frac{C_{Effluent}}{C_{Influent}} \right) \quad (14)$$

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4 300 The pC^* parameter is a convenient dimensionless measure of particle removal efficiency. In
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6 301 this study, pC^* was calculated for overall removal efficiency of the floc blanket and tube set-
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8 302 tler ($pC^*_{Overall}$), for the floc blanket ($pC^*_{FlocBlanket}$), and for the tube settler ($pC^*_{TubeSettler}$).
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10 303 By definition, $pC^*_{Overall}$ is the sum of $pC^*_{TubeSettler}$ and $pC^*_{FlocBlanket}$.
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12 304

13 14 15 305 RESULTS AND DISCUSSION

16 17 306 *Replicability and stability of the bench-scale apparatus*

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19 307 Four replicate experiments were carried out to confirm that a floc blanket could be
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21 308 formed consistently in the upflow clarifier, and to obtain baseline performance data once
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23 309 steady-state was reached. Achieving consistent performance in these control experiments
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25 310 enhanced the ability to identify cases where floc roll-up caused an elevated effluent turbidity.
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27 311 Each replicate experiment used a 25.4 mm diameter tube settler downstream of the upflow
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29 312 clarifier. Two of the four trials were performed using a reservoir to sample the tube settler
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31 313 effluent turbidity. Exemplary results from one of these four tests are shown in Figure 5.
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33 314 Region A in the figure shows the period of floc blanket formation, before the system reaches
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35 315 steady-state performance in Region B.

36
37 316 The results indicated that the bench-scale system could produce replicable data. Once
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39 317 the system reached steady state, good baselines were observed for all measured turbidity
40
41 318 values as shown in Figure 5. For four replicate trials, the raw water was maintained at
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43 319 around 101.4 NTU with an average coefficient of variation of 5%, and the clarified effluent
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45 320 was consistently around 11.4 NTU \pm 13% after the floc blanket had formed. The tube
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47 321 settler effluent was 0.22 NTU \pm 43%; however, this higher variability was considered accept-
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49 322 able given the small magnitude of the effluent turbidity. The tube settler effluent data from
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51 323 the reservoir trials was consistent with the data from the non-reservoir trials, indicating that
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53 324 the sampling reservoir did not introduce a systematic bias into the tube settler turbidity data.
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55 325

326 *Floc roll-up experiments*

327 Table 1 summarizes the experimental conditions used to investigate floc roll-up. The
328 tube settlers were sized using Equation (1) to maintain a constant settle capture velocity
329 of 0.10 mm/s. The flow rate through the tube settler was controlled by a peristaltic pump.
330 This flow rate Q was calculated from:

$$331 \quad Q = V_{Settle} \left(\frac{L}{D} \cos \alpha + \sin \alpha \right) \pi \frac{D^2}{4} \quad (15)$$

332 where D is the inner diameter of the tube, V_{Settle} is the settle capture velocity set at 0.10
333 mm/s, and α is the angle of inclination set at 60°. Because the settle capture velocity was
334 held constant, each tube settler described in the table would be expected to have identical
335 turbidity removal performance based on conventional design. Variations in performance
336 therefore would show the effect of the floc roll-up failure mode on particles that would
337 otherwise have been expected to be captured.

338 Figure 6 gives the measured pC^* as a function of velocity gradient at the tube wall for
339 the experiments listed in Table 1. When velocity gradient increased, the performance of the
340 system declined, while the performance of the floc blanket remained relatively consistent
341 with an average pC^* of 1.12. At low velocity gradients, the overall system had a pC^* of
342 2 or greater, but this declined to just over 1 for higher velocity gradients. This effect was
343 attributable to the tube settler, because there was no systematic variation in the particle-
344 removal performance of the floc blanket.

346 *Comparison to theoretical predictions*

347 The results show a decrease in tube settler performance with increasing velocity gradient,
348 despite the fact that each experiment had a settle capture velocity of 0.10 mm/s. The
349 experimental observations are qualitatively consistent with model predictions - high velocity
350 gradients were expected to cause flocs to roll up the inclined surface and thus increase effluent

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4 351 turbidity. Under the experimental conditions, the slide capture velocity usually controlled
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6 352 tube settler performance, which led to the variation observed in Figure 6 even though each
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8 353 tube settler had the same settle capture velocity.

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10 354 Figure 7 shows the extent to which the roll-up phenomenon affected the tube settler in
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12 355 each experiment. The observed turbidity removal performance of the tube settler (as pC^*) is
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14 356 plotted against the ratio of the slide capture velocity V_{Slide} (estimated from Equation (10))
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16 357 to the settle capture velocity V_{Settle} (fixed at 0.10 mm/s). The higher this ratio, the poorer
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18 358 the performance of the tube settlers, which suggests that the floc roll-up failure mode was
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20 359 indeed responsible for the decline in performance as velocity gradient increased. Note in the
21
22 360 graph that the tube settler achieved good and relatively consistent performance until the
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24 361 capture velocity ratio reached a value of 1. Once the slide capture velocity was greater than
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26 362 the settle capture velocity, performance declined as more particles began to escape via floc
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28 363 roll-up.

31 32 33 364 34 35 365 *Practical implications*

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37 366 The results presented above, along with the theoretical model proposed in Equation
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39 367 (10), suggest that the typical spacing of inclined settling devices can indeed be reduced. The
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41 368 following example depicts the possible material savings on sedimentation tanks by halving
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43 369 plate settler spacing. The length of plate settlers is typically about 20 times as long as
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45 370 the spacing. Conventional design guidelines suggest a spacing of 5 cm, resulting in plate
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47 371 lengths of approximately 1 m. Assuming an angle of inclination of 60°, these plate settlers
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49 372 would occupy 0.86 m of sedimentation tank depth. Reducing the spacing to 2.5 cm would
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51 373 reduce the required sedimentation tank depth by 0.43 m and significantly lower the cost of
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53 374 construction.

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55 375 Based on this insight, the AguaClara program at Cornell University has been using a
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57 376 design spacing of 2.5 cm for inclined plate settlers in upflow sedimentation tanks, for the
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59 377 affordable gravity-fed drinking water treatment plants it has implemented in seven towns in
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4 378 Honduras. These plate settlers are designed for a capture velocity of 0.12 mm/s, and they
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6 379 perform well even with the reduced plate settler spacing and reduced tank depth (Smith,
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8 380 2010).

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10 381 For any water treatment application, the slide capture velocity could be calculated, and
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12 382 the spacing of the inclined settling device could then be selected to ensure that the slide
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14 383 capture velocity does not exceed the settle capture velocity. The minimum allowable spacing
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16 384 for plate settlers can be obtained by setting the settle capture velocity (Equation (1)) equal
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18 385 to the slide capture velocity (Equation (10)) and solving for the spacing:

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$$D_{Min} \approx \frac{3}{\sin^2 \alpha} \frac{V_{\uparrow}}{V_C} d_{Floc0} \left(\frac{18V_{\uparrow} \Phi \nu_{Water}}{gd_{Floc0}^2} \frac{\rho_{Water}}{\rho_{Floc0} - \rho_{Water}} \right)^{\frac{1}{D_{Fractal}-1}} \quad (16)$$

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24
25 387 where $V_C = V_{Settle} = V_{Slide}$ and the constant 3 is used for the plate geometry rather than
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27 388 the 4 that would be used for tube geometry. Equation (16) is based on first principles and
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29 389 provides a basis for evaluating the influence of various parameters on the required spacing.
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31 390 The calculated spacing is the minimum value required to avoid the adverse effect of floc
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33 391 roll-up, based on the observation that performance degradation begins to occur when the
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35 392 slide capture velocity is greater than the settle capture velocity. The required spacing is
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37 393 a function of floc properties (primary particle density - ρ_{Floc0} , fractal dimension - $D_{Fractal}$,
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39 394 primary particle diameter - d_{Floc0}), water properties (viscosity - ν_{Water} and density - ρ_{Water}),
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41 395 upflow velocity - V_{\uparrow} , and desired capture velocity - V_C . Equation (16) is plotted in Figure
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43 396 8 for a base case of 20° C water, 100 mg/L of clay particles with a diameter of 2 μm , plate
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45 397 settlers angled at 60°, capture velocity of 0.12 mm/s, and a fractal dimension of 2.3. Three
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47 398 other cumulative effect cases are also shown in Figure 8. In the second case, the clay is
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49 399 removed from the raw water and the primary particle is a 100 nm aluminum hydroxide pre-
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52 400 cipitate. In the third case, the viscosity of water is also increased to that of 0° C water. In
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54 401 the fourth case, the fractal dimension is also decreased to 2.2. This analysis suggests that
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56 402 plate settler spacing could be much smaller for clay-alum flocs, but that larger spacing may
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58 403 be required for low-temperature water with flocs formed from smaller-size primary particles.
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4 404 Although the fractal dimension is expected to be relatively constant for flocs that are formed
5 405 under the same aggregation conditions, the dramatic impact of a small change in the fractal
6 406 dimension suggests that further work to characterize the fractal dimension of flocs formed
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8 407 at water treatment plants would be beneficial.
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13 14 15 409 **CONCLUSIONS**

16
17 410 The floc roll-up model describes a failure mechanism that prevents flocs from sliding
18 411 along an inclined surface in the counter-current direction. This failure is caused by velocity
19 412 gradients at the plate or tube wall, which create a fluid drag on the floc that opposes grav-
20 413 itational forces. A theoretical model is presented as an analysis of this failure mechanism,
21 414 leading to a “slide capture velocity” for situations in which floc roll-up controls the perfor-
22 415 mance of inclined settling devices. Consistent with model predictions, experimental inclined
23 416 tube settlers showed a decline in performance with increasing velocity gradient, even though
24 417 traditional design equations predict that the settle capture velocity and performance of these
25 418 tube settlers should have been constant. Tube settler turbidity removal deteriorated when
26 419 the slide capture velocity was larger than the settle capture velocity. This failure mode ex-
27 420 plains the rationale for setting a minimum value for the spacing of inclined settling devices,
28 421 but it appears that there are opportunities to use smaller spacings to create more economical
29 422 sedimentation tanks. Further work to characterize the properties of flocs from a variety of
30 423 water sources, coagulant dosages, and types of coagulants would be helpful to determine the
31 424 limiting case for plate settler spacing or tube settler diameter.
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49 50 51 426 **ACKNOWLEDGMENTS**

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12 434 the Planet” design competition.

435

16 17 436 **AUTHOR DISCLOSURE STATEMENT**

18
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TABLE 1. CONFIGURATION OF THE FLOC ROLL-UP EXPERIMENTS.

<i>Tube Geometry</i>		<i>Flow Rate</i>	<i>Incline Velocity</i>	<i>Velocity gradient</i>
<i>ID (mm)</i>	<i>Length (m)</i>	<i>(mL/min)</i>	<i>V_a (mm/s)</i>	<i>at tube wall (s⁻¹)</i>
6.35	0.12	1.90	1.00	1.26
6.35	0.24	3.79	1.99	2.51
6.35	0.40	6.18	3.25	4.10
6.35	0.62	9.49	4.99	6.29
6.35	0.65	14.44	7.60	9.57
6.35	1.15	17.37	9.14	11.52
6.35	1.20	18.18	9.57	12.06
6.35	1.83	27.53	14.49	18.25
9.53	0.36	8.54	2.00	1.68
9.53	0.93	21.34	4.99	4.19

Velocity gradient at tube wall calculated with Equation (3).
ID, inner diameter.

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3 **FIG. 1.** Schematic of the system geometry and conditions experienced by a floc on the bottom
4 surface of a tube or plate settler.
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10 **FIG. 2.** Theoretical tube settler slide capture velocities (V_{Slide}) for tubes of three different inner
11 diameters. Here, the slide capture velocity is the minimum terminal settling velocity of a floc that
12 will be retained on a surface inclined at $\alpha = 60^\circ$ because of a balance between gravity forces and
13 fluid drag. The floc parameters assumed for this calculation were: $D_{Fractal} = 2.3$, $d_{Floc0} = 1 \mu\text{m}$, Φ
14 $= 45/24$, and $\rho_{Floc0} = 2624 \text{ kg/m}^3$.
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24 **FIG. 3.** Apparatus to prepare influent water for the experimental treatment system, with aeration,
25 temperature control, clay addition, and influent turbidity sampling.
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32 **FIG. 4.** Bench-scale water treatment apparatus, including an alum dosing pump, a rapid mix
33 tube, a hydraulic tube flocculator, a sedimentation column with a floc blanket, and a tube settler.
34 Turbidity was monitored above the floc blanket and after the tube settler.
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41 **FIG. 5.** Example of influent turbidity, floc blanket effluent, and tube settler effluent data from a
42 control experiment. Region A shows the time interval for floc blanket formation, and Region B
43 shows subsequent steady-state performance.
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50 **FIG. 6.** Log turbidity removal pC^* for the overall treatment system and for the floc blanket, as a
51 function of the velocity gradient at the wall of the tube settler.
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3 **FIG. 7.** The effect of the ratio of V_{Slide} to V_{Settle} on tube settler performance. Performance
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5 deteriorates when V_{Slide} is greater than V_{Settle} .
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10 **FIG. 8.** Minimum spacing for plate settlers to avoid floc roll-up, calculated as a function of
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12 upflow velocity under four different scenarios.
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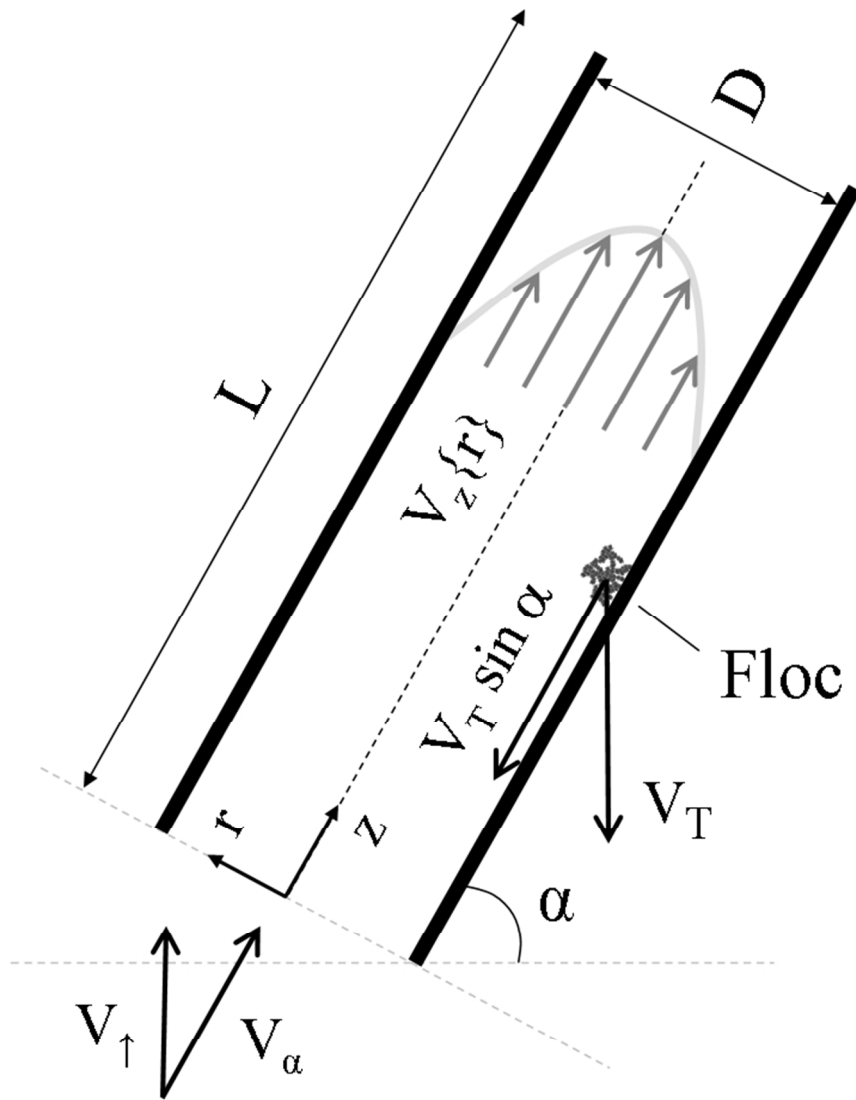


FIG. 1. Schematic of the system geometry and conditions experienced by a floc on the bottom surface of a tube or plate settler.
141x165mm (150 x 150 DPI)

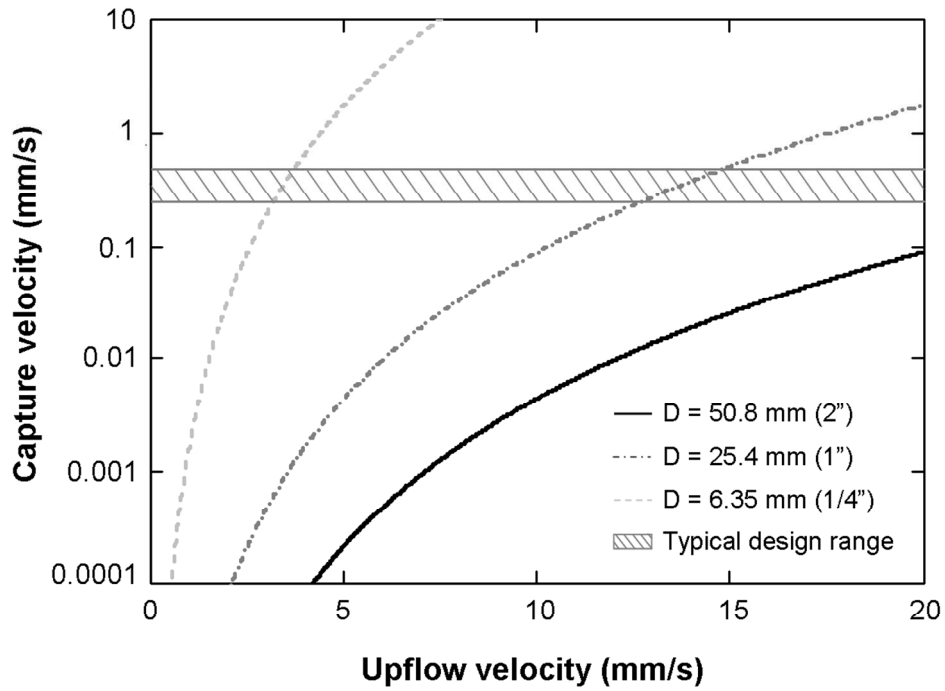


FIG. 2. Theoretical tube settler slide capture velocities for tubes of three different inner diameters.
224x165mm (150 x 150 DPI)

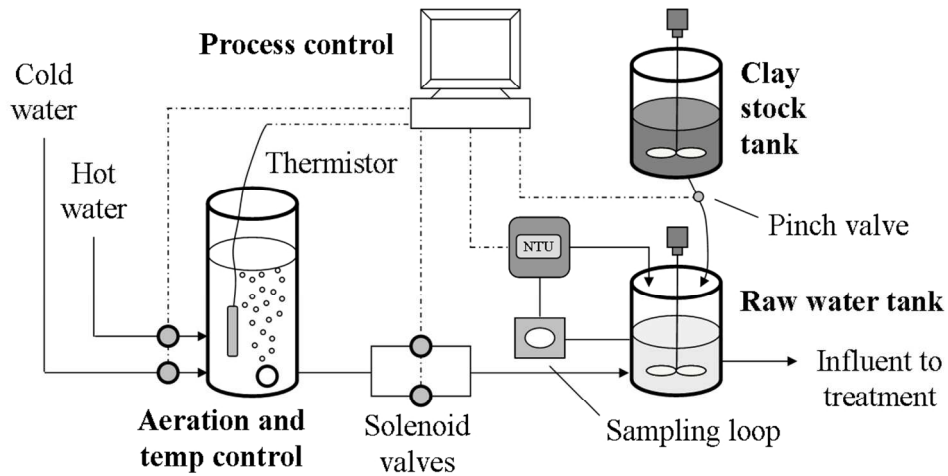


FIG 3. Apparatus to prepare influent water for the experimental treatment system, with aeration, temperature control, clay addition, and influent turbidity sampling.
213x112mm (150 x 150 DPI)

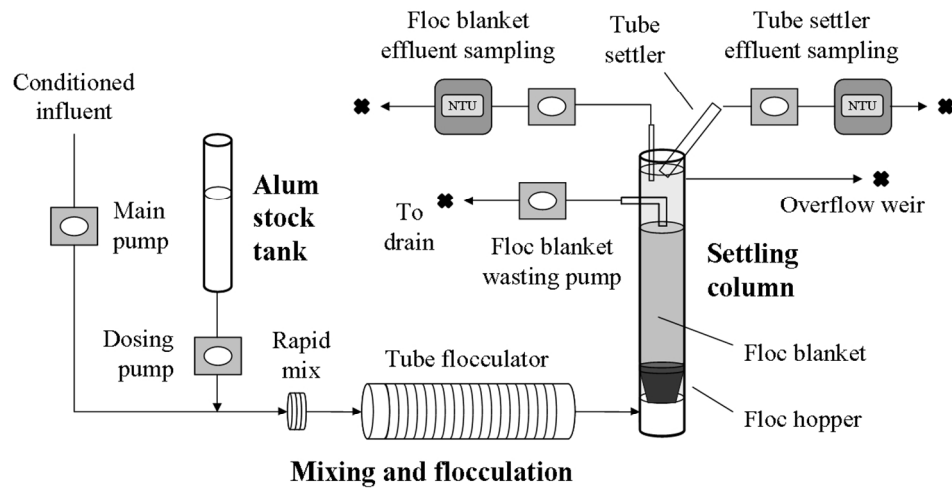


FIG. 4. Bench-scale water treatment apparatus, including an alum dosing pump, a rapid mix tube, a hydraulic tube flocculator, a sedimentation column with a floc blanket, and a tube settler.
228x122mm (150 x 150 DPI)

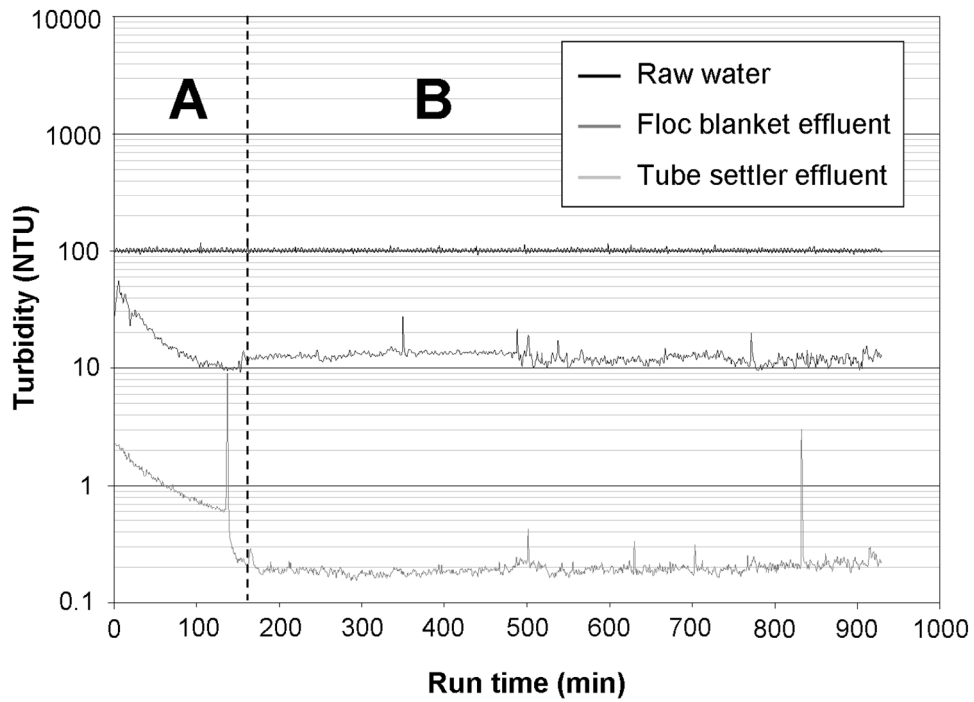


FIG. 5. Example of influent turbidity, floc blanket effluent, and tube settler effluent data from a control experiment.
249x182mm (150 x 150 DPI)

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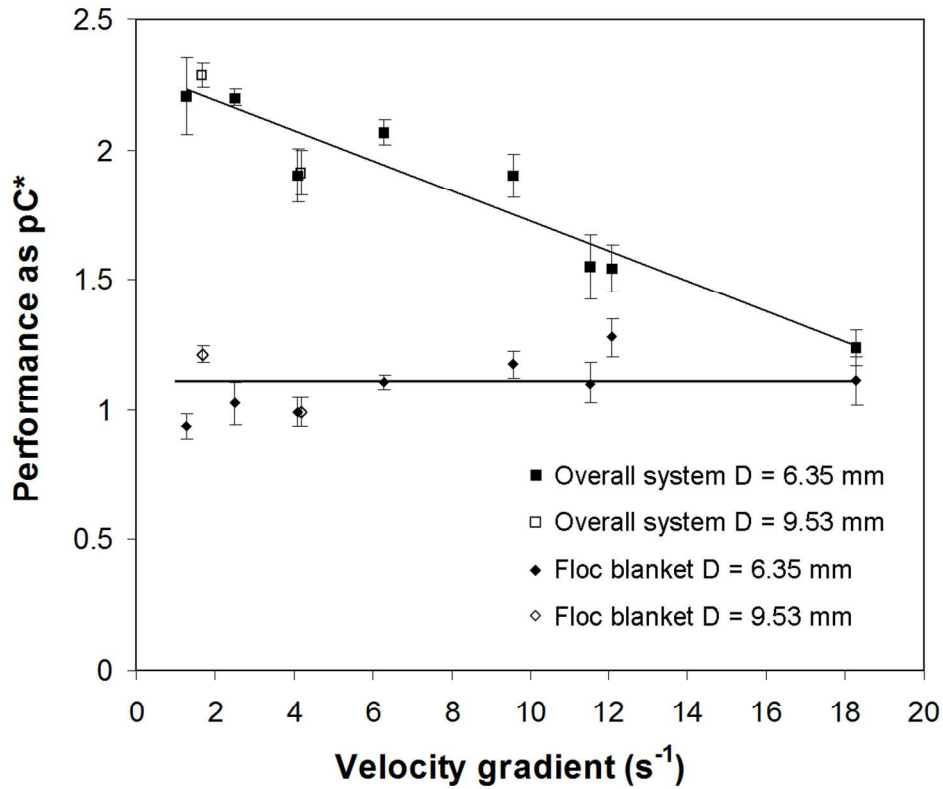


FIG. 6. Log turbidity removal pC* for the overall treatment system and for the floc blanket, as a function of the velocity gradient at the wall of the tube settler.
215x177mm (150 x 150 DPI)

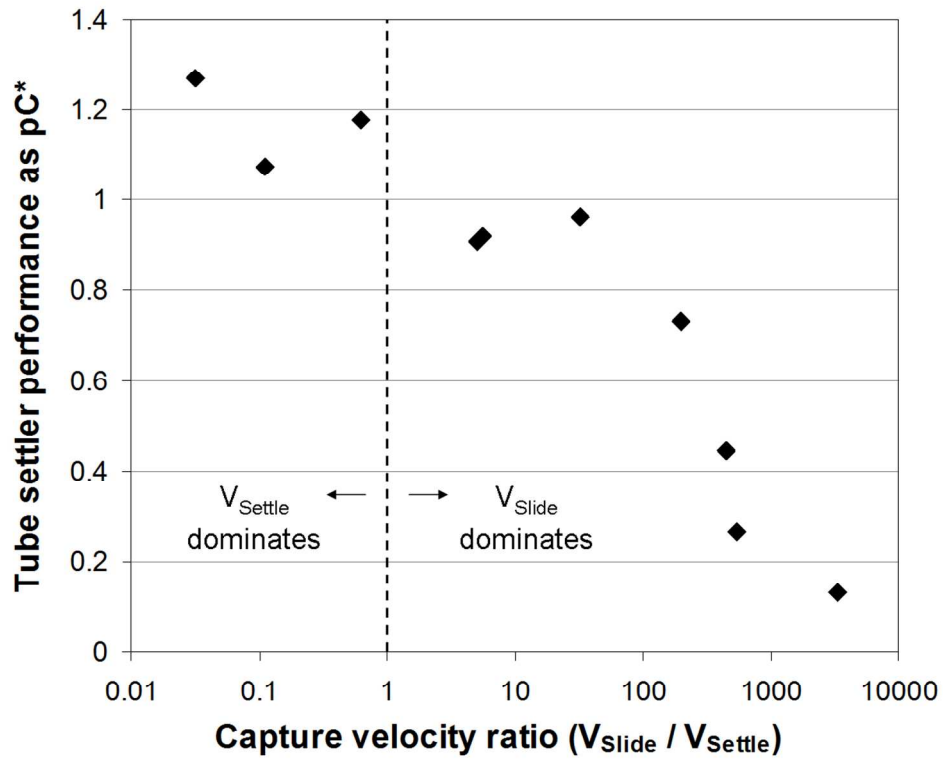


FIG. 7. The effect of the ratio of V_{Slide} to V_{Settle} on tube settler performance.
217x178mm (150 x 150 DPI)

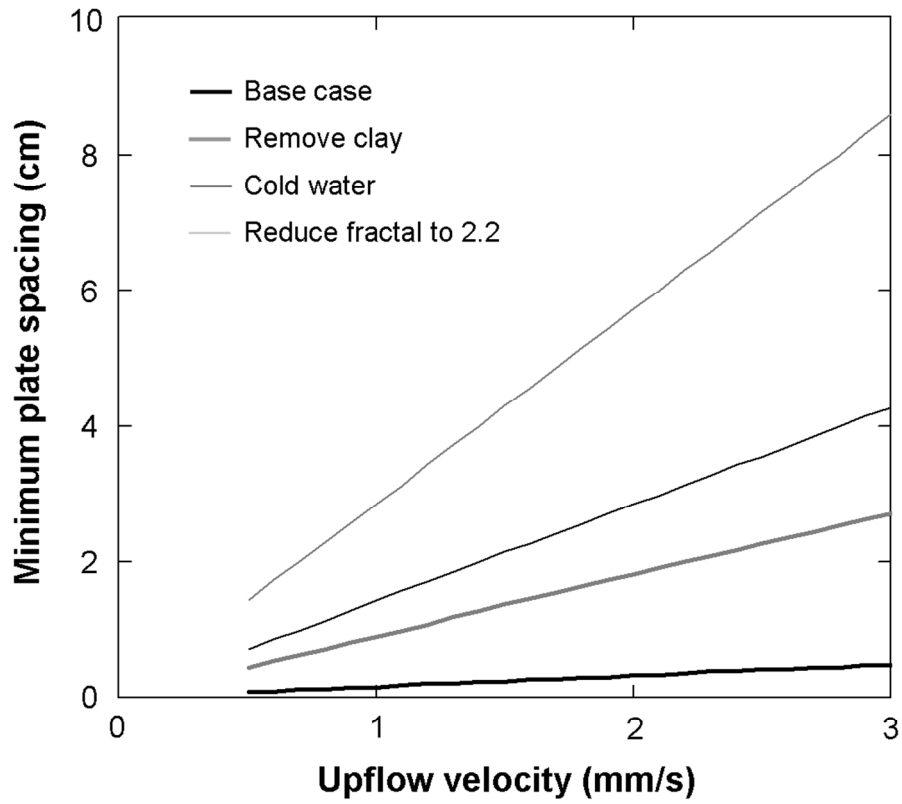


FIG. 8. Minimum spacing for plate settlers to avoid floc roll-up, calculated as a function of upflow velocity under four different scenarios.
203x173mm (150 x 150 DPI)